

HOW CAN NEURAL NETWORKS SPEED UP ECOLOGICAL REGIONALIZATION FRIENDLY?

Replacement of Field Studies by Satellite Data using RBFs

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Abstract: The aim of this work is to present an application of the Radial Basis Functions Nets (RBFs) for simplifying and reducing the cost of ecological regionalization. The process speeds up and replaces the classic means of obtaining ecological variables through field studies. The radial basis function networks were applied to estimate field data remotely, using data captured by the Landsat satellite and correlating it with ecological variables in order to substitute for them in the regionalization process. This approach substantially reduces the time and cost of ecological regionalization, limiting field studies and automating the generation of the ecological variables. The technique could be applied without restriction to map vegetation in any other area for which satellite coverage exists.

1 INTRODUCTION

The need or sound environmental management of a particular territory requires a sufficient and integrated understanding of the resources there, and the interrelationships with the natural and human elements that act upon it (Moreira, 2000). Nevertheless, and in spite of the enormous effort to generate this thematic information, the reality is that the results are still unsatisfactory in terms of environmental planning (Pablo, 2000).

This situation has led the majority of the Spanish regions to prepare their own environmental cartographic information. Ecological regionalization was developed with the aim of providing useful information about the web of relationships between various natural elements in an area over both space and time. This type of mapping attempts to integrate the most relevant environmental aspects of a territory in order to identify patterns that allow the structure and operation of a territory to be understood, classifying them into a series of units called *environmentals*. These units are characterized by the close homogeneity of the environmental variables considered (Naiman et al., 1992). Figure 1 shows the vegetation

map based on the ecological variables used for this study.

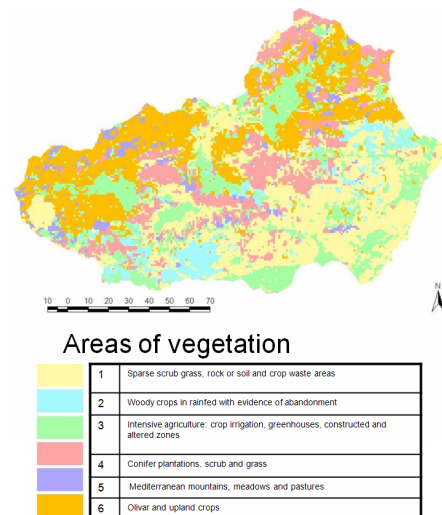


Figure 1: Vegetation map based on the ecological variables used in this study area (Almería and Granada, Spain).

Ecological regionalizations (also referred to as ecoregions, biogeographic regions land classes and

environmental classifications, environmental domains, ecological land units,) play important roles in process of resource management by grouping locations with similar environmental variables (Snelder et al., 2010). Undertaking this kind of ecological regionalization requires the following environmental data (Loveland and Merchant, 2004):

- Climatic: obtained from the state meteorological agencies, and comprising time series running over several years.
- Geological: obtained from field studies to measure rainfall and soil composition.
- Vegetation cover: obtained from field studies of vegetation cover in order to tabulate the presence of different species and vegetation types.
- Land Use: obtained from field studies of land use in the study area.

This information is used by the various technical environmental departments for ecological regionalization, using statistical techniques. Nonetheless, until now, it has still been necessary to undertake various field studies to keep the environmental information up-to-date, and this makes it quite difficult to have an up-to-date ecological map in a short period of time. Thus, with the dual aims of unifying the information sources used to draw ecological maps, and simplifying and reducing the cost of obtaining many of these data, the present study (undertaken within the framework of the SOLERES project financed by the Spanish Ministry of Science and Technology) applies models of neural networks that use cheaper and up-to-date satellite data to construct ecological maps, whilst maintaining the procedures for constructing this cartography. The study is not designed to substitute one set of ecological variables for another, but to find an alternative procedure to construct the variables.

2 NEURAL NETS AND REMOTE SENSING

Neural networks have proved their potential as tools for classification and approximation of functions for over ten years. The advantages over other analytical and statistical techniques stems from their capacity to handle large volumes of high-dimensional data, the capacity to work with scarce, changing and/or contradictory data, and their independence from the statistical characteristics of the sample. Moreover, neural networks exhibit a significant capacity for generalisation that means they become powerful tools for studying the dynamics of atmosphere and climate from

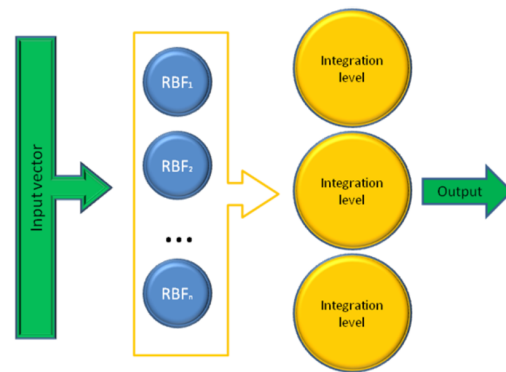


Figure 2: Basic structure of an RBF.

satellite images, for classifying vegetation types and other related activities (Richards, 1993).

Radial basis function networks (RBFs) (Poggio and Girosi, 1990) were developed later than MLPs and, from an operational point of view, have certain advantages over MLPs, such as faster training and an internal structure that allows better understanding of the relationships between variables, and therefore, a better understanding of how the network functions. Figure 2 shows the three basic levels of an RBF:

- An input layer, in which the characteristics vector is applied to each and every element of the following level.
- A radial basis function net level, called the hidden layer unit, which computes the expression:

$$RBF_i(\vec{x}) = e^{-\frac{\|\vec{x}-\vec{c}_i\|^2}{\sigma^2}}$$

where c_i is the centroid of the radial base function and σ is the scope parameter (measuring the spread) of each radial base function.

- An integration or output layer, where the results of each RBF is adjusted/weighted, to give the output of the net, according to the function:

$$Y = \sum_i w_i RBF_i(\vec{x}) - b$$

where w_i is the weighting parameter and b is the threshold value.

The training phase of an RBF consists of selecting the centroids c_i and the values of the weights w_i that minimise the error produced by the network for the training data set. RBFs exhibit certain advantages over other types of net like MLPs, such as the speed of training. Nevertheless, they have certain drawbacks, such as the fact that, to solve a particular problem, RBFs require a greater number of neurons and so a larger computing effort. The application of RBFs for the present study is justified by the fact that

many training phases are required and this would take a long time using other methods.

3 DESCRIPTION OF THE WORK

The study area comprised the provinces of Almería and Granada, in south-eastern Spain (southern Europe). The study was designed in several phases:

1. Determine the ecological variables that are suitable for approximation using satellite data.
2. Obtain satellite data to correspond with each value of the ecological variable.
3. Train and run simulations of the neural networks.
4. Compare the results obtained with the expected values.

3.1 Determine the Ecological Variables that are Suitable for Approximation using Satellite Data

The data for these variables were obtained from field studies undertaken by technical staff working for the Administration. The data are updated periodically using field data or aerial photographs. The information for each variable is obtained by weighting the surface area of vegetation cover for each vegetation type within each 1x1Km sector. This produces numerical values for each 1x1Km sector and for each variable, which represents the percentage cover of each vegetation type included in this study, expressed over the interval [0,100].

The area of study includes a great diversity of woodland landscapes, ranging from wet woodlands to desert landscapes containing a wide range of ecosystems, dominated by high mountain ranges, subtropical maritime area and subdesert plains. In this way, the study enabled an analysis of landscapes containing a wide variety of vegetation types (Snelder et al., 2007). Table 1 shows the ecological variables used.

3.2 Obtain Satellite Data to Correspond with Values of Ecological Variables

Improvements in remote sensing technologies and the use of geographic information system (GIS), are increasingly allowing us to develop indicators that can be used to monitor and assess ecosystem condition and change at multiple scales (Revenga, 2005). The Landsat satellite includes a TM sensor (Thematic Mapper) that captures data over seven bands of the

electromagnetic spectrum. The sensor is linked to territorial studies whose principal focus is environmental. Its orbit, synchronous with the sun, is at an altitude of 705km and has a period of 98.9 minutes, totalling 14 flights daily around the earth.

Table 1: List of ecological variables used in the study.

N	Concept
01	Built-up and disturbed areas
02	Wetlands and open water
03	River corridors vegetation
04	Herbaceous crops without irrigation
05	Olive plantations
06	Other woody crops without irrigation
07	Forced crops under plastic
08	Other herbaceous crops under irrigation
09	Woody crops under irrigation
10	Herbaceous and woody crops
11	Mosaic of woody crops (no irrigation)
12	Herbaceous and woody crops (irrigation)
13	Mosaic of woody crops under irrigation
14	Mosaic of crops under irrigation or not
15	Herbaceous crops and pastures
16	Herbaceous crops and woody vegetation
17	Woody crops and pastures
18	Woody crops and woody vegetation
19	Mosaics of crops and natural vegetation
20	Abandoned woody crops
21	Mediterranean oak woodland
22	Conifers
23	Other tree species and mixtures
24	Dense matorral + dense oak wood
25	Dense matorral + scattered oaks
26	Dense matorral + dense conifers
27	Dense matorral + scattered conifers
28	Dense matorral + mixed tree species
29	Scattered matorral + dense oak wood
30	Scattered matorral and oaks
31	Scattered matorral + dense conifers
32	Scattered matorral + scattered conifers
33	Scattered matorral + mixed tree species
34	Pasture + dense Mediterranean oakwood
35	Pasture + scattered Mediterranean oaks
36	Pasture + dense conifers
37	Pasture + scattered conifers
38	Pasture + other tree species or mixed
39	Zones subject to fierce erosion or fires
40	Dense matorral
41	Scattered matorral with pasture
42	Scattered matorral (pasture and rock)
43	Continuous pasture
44	Pasture with rock or soil clearings
45	Beaches, dunes and sandbanks

In order to capture all of the study area, we needed three Landsat scenes. These satellite images were used to make a mosaic, and the image was trimmed with a vector of the outline of Almería and Granada provinces, giving the processed image that served as the basis for our study (vegetation data and remote sensing matching).

To do this, the original image was trimmed using a window of 33x33 pixels (990m²). The displacement error (the 10m² missing from the 1000m²) was corrected by alternating the 33x33 pixel window with another measuring 34x34 pixels, every two passes. The result was an image of 241x157 pixels, with a spatial resolution of 1000x1000m.

Following the same procedure, the median, mean and 30 percentile of each window of the original image was also calculated. This was stored, for each of its seven layers, in a matrix. By this means, we generated 21 variables with which to calculate each of the ecological variables of vegetation cover. Figure 3 shows Almería and Granada provinces in LandSat data.

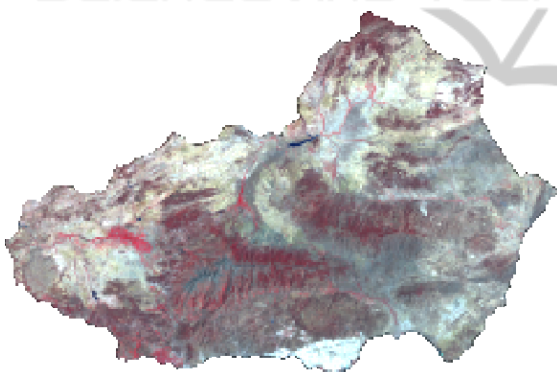


Figure 3: Detail of the study area using LandSat data.

3.3 Train and Run Simulations of the Neural Networks

The next step was to train a series of RBF neural networks to relate each ecological variable to the processed satellite data. For each ecological variable, there are 21,905 ground surface samples available, each representing 1x1 Km, for which statistical information from the seven satellite bands has been generated. In this way, each sector yields 21 characteristics describing it. So, for each variable we have a data model of 21,905x21, with 21,905 desired results, which are the data from which we will train the neural net.

Modelling of each of the 45 ecological variables involved repeating the construction of an RBF net 32 times, using 70% of the data as the training data set

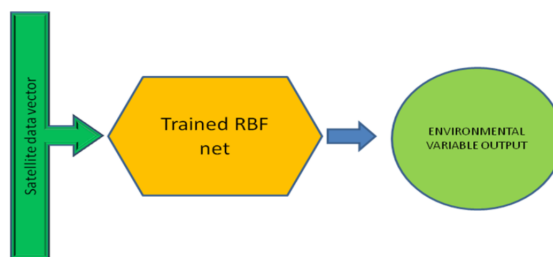


Figure 4: Detail of the process undertaken by the neural net.

and the remaining 30% as the calibration data set. The division of data into these two datasets was done at random in every case. In total, 1440 networks were trained.

Once trained, the input to the neural net is changed to the satellite data, and its output offers an approximate value of the environmental variable for which it has been trained. This technique increases the possibility of finding a net that satisfactorily approximates to the ecological variable. The radial basis function network uses a random parameter over the interval [0, 1.5]. Figure 4 shows the process undertaken by the neural net.

3.4 Compare the Results obtained with the Expected Values

Once the networks are constructed, the results obtained for the calibration data set are compared with the expected values and two measures are calculated:

1. The *precision*, obtained using the formula:

$$pr = \frac{e}{|C|}$$

where e is the number of times that the output of the neural net coincides with the expected output, with an error of ± 0.005 , and C is the calibration data set. The training data set is not taken into consideration because the precision obtained with the training set would be close to 1 and this would distort the precision of the calibration data set.

2. The *variance* between the expected results and those obtained indicates the ecological variables for which confidence in the approximation is sufficiently high. To calculate variance, both expected and obtained results are normalised to a range [0,1] and the variance of the difference in the absolute values is calculated.

The process was programmed in Matlab, using a desktop computer with 4 core architecture and 8 GB of memory. The program took 48 hours of intensive calculations. An improvement in the calculation

time would be achieved using other computer architectures.

4 RESULTS

The most obvious finding of the experiment is that the result was positive for all the ecological variables, except variable 34 (pasture and dense Mediterranean oak woodland) and variable 37 (pasture with scattered conifers). The neural networks obtained fit the behaviour of the ecological variables with a precision of 0.8 and variance of less than 0.03 in every case. For the remaining two ecological variables that could not be simulated, had given training networks with a peak precision of 0.70 for the first and 0.42 in the second. Figure 5 shows the distribution of the experiment precision by variable.

A priori, the result is not explained by the nature of the variable and we find the explanation in the procedure used to choose the training and calibration sets: the choice is made at random and with a relatively small frequency of non-null data, the distribution between the training and calibration data sets is significant for the precision of the net. Thus, for example, if we use all the non-null data in the training phase, we will obtain a precision of 0.9998 and a variance of 0.0001 for variable 34; and a precision of 0.9983 and variance of 0.0016 for variable 37. Table 2 shows the relative frequency of non-null data by variable.

Excluding variables 34 and 37, the number of experiments that generated networks of an acceptable precision was, on average, 68%. 70% of the variables developed at least 53% precise neural networks. This finding permits discussion of whether the model of fitting ecological variable using Landsat data and RBF networks is sound. A significant number of experiments were undertaken for each variable, varying the training data set and this achieved good results, in the majority of cases, improving as the quantity of non-null data increases for the ecological variable in question.

With respect to the parameters of the network, within the interval [0,1.5] the parameter scope does not appear to be a determining factor in achieving a better result. It is relevant to point out that, in a significant proportion of cases, the neural networks did not manage to approximate to the associated ecological variable. This situation is due to the nature of the data for each variable and the randomness of selecting data during each training process.

From an environmental point of view, organization of the data into ecological variables and its subse-

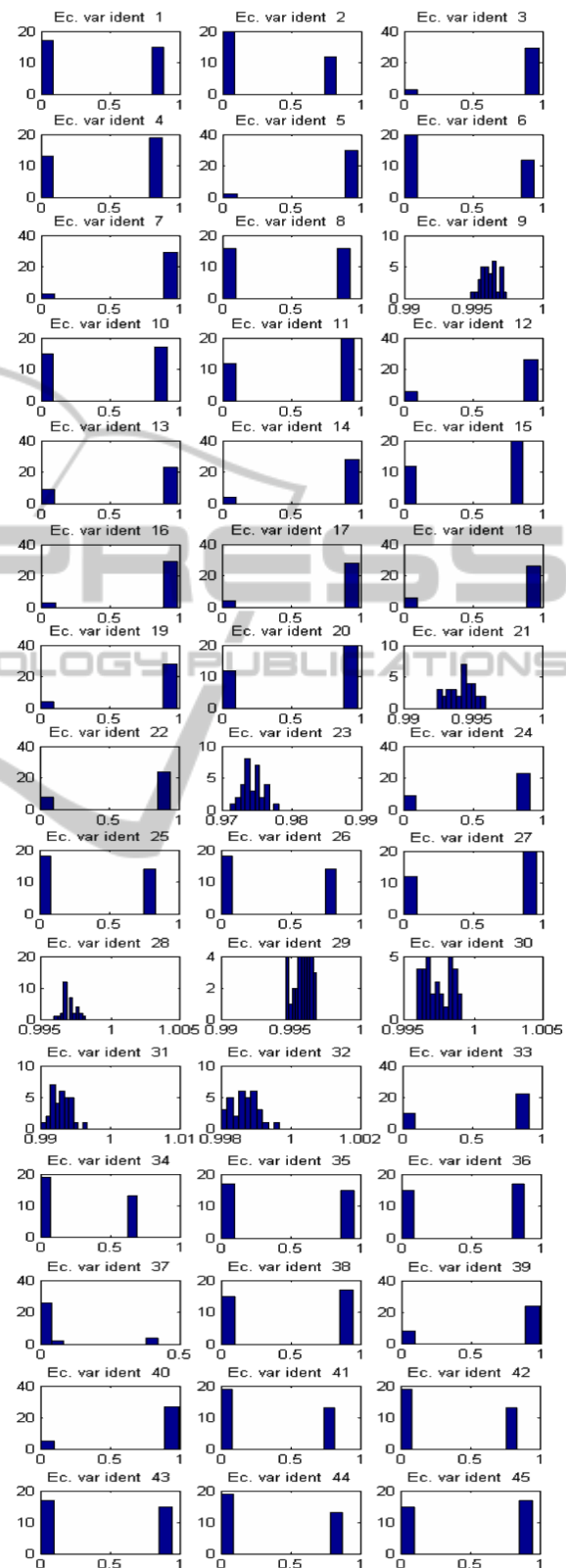


Figure 5: Distribution of the experiment precision by variable.

Table 2: Relative frequency of non-null data.

id 01	id 02	id 03	id 04	id 05
0,115	0,036	0,024	0,308	0,179
id 06	id 07	id 08	id 09	id 10
0,164	0,052	0,129	0,057	0,176
id 11	id 12	id 13	id 14	id 15
0,030	0,128	0,021	0,061	0,013
id 16	id 17	id 18	id 19	id 20
0,083	0,004	0,093	0,049	0,014
id 21	id 22	id 23	id 24	id 25
0,017	0,145	0,015	0,012	0,018
id 26	id 27	id 28	id 29	id 30
0,015	0,024	0,006	0,066	0,097
id 31	id 32	id 33	id 34	id 35
0,166	0,174	0,046	0,003	0,004
id 36	id 37	id 38	id 39	id 40
0,002	0,006	0,001	0,091	0,048
id 41	id 42	id 43	id 44	id 45
0,116	0,583	0,049	0,011	0,006

quent substitution using satellite data can be successfully achieved, as proved by the experience in the classification of vegetation types using the LandSat data.

In addition, at least in the study zone, there is no noticeable interaction between various vegetation covers to complicate the training of the networks which, in other situations, would have to be studied. This situation may be due in part to the values obtained: for each quadrat of land, the values indicate a dominant vegetation type, so that 90% of the samples have a vegetation cover of more than 41.9, while 50% of cases have an vegetation cover exceeding 68.8. In such a situation, the information obtained from the satellite data is representing, in the majority of cases, the dominant characteristic and for this reason, there are no undesired interactions. In addition, the matrix contains a large number of nil or very low data values: 92.3% of data are zero and only 55,800 data of a total of 985,725 in the matrix, have values above 5.

5 CONCLUSIONS

As a final point, the approximation functions of the ecological variables developed here using radial basis function networks could be used in subsequent years to study changes in vegetation cover. Although the vegetation cover changes seasonally, it is also true that the experiment could be repeated for different seasons of the year, so long as this cover existed.

The use of Landsat data in this case reduces field studies in at least 30%. Neural networks can recognize geographical locations with similar vegetation characteristics at any given time. This situation will allow work teams to study the Landsat information

previously available and improve the work on a surface, saving costs.

From a technical point of view, the study also corroborates the need for a precise study of the training data set in order to achieve a precise training so that the results are consistent with the environmental model simulated. The results confirm our working hypothesis that supports the viability of a computation process of ecological variables that uses satellite data that could substitute for the traditional field studies.

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REFERENCES

- Loveland, T. R. and Merchant, J. M. (2004). Ecoregions and ecoregionalization: geographical and ecological perspectives. In *Environmental Management 34*. Springer New York.
- Moreira, J. (2000). *Reconocimiento biofísico de espacios naturales protegidos. Parque natural Sierras Subbéticas*. Junta de Andalucía, Sevilla, 1st edition.
- Naiman, R., Loranrich, D., Beechie, T., and Ralph, S. (1992). General principles of classification and the assessment of conservation potential in rivers. In *River Conservation and Management*. John Wiley and Sons.
- Pablo, C. D. (2000). Cartografía ecológica: conceptos y procedimientos para la representación espacial de ecosistemas. In *Boletín de la Real Sociedad Española de Historia Natural*. Real Sociedad Española de Historia Natural.
- Poggio, T. and Girosi, F. (1990). A theory of networks for approximation and learning. In *Proceedings of the IEEE 78*. Massachusetts Institute of Technology.
- Revenga, C. (2005). Developing indicators of ecosystem condition using geographic information systems and remote sensing. In *Regional Environmental Change 5*. Springer Berlin and Heidelberg.
- Richards, J. (1993). *Remote sensing digital image analysis. An introduction*. Springer-Verlag, Berlin, 2nd edition.
- Snelder, T., Leathwick, J., and Dey, K. (2007). A procedure for making optimal selection of input variables for multivariate environmental classifications. In *Conservation Biology 21*. National Institute of Water and Atmospheric Research.
- Snelder, T., Lehmann, A., Lamouroux, N., Leathwick, J., and Allenbach, K. (2010). Effect of classification procedure on the performance of numerically defined ecological regions. In *Environmental Management 45*. Springer New York.